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Butterfly richness and abundance in flower strips and field margins: the role of local habitat quality and landscape context



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ABSTRACT

Flower strips, which are created on arable land by sowing species-rich seed mixtures, are considered to have a high potential to counteract species decline of butterflies in the agricultural landscape. However, it remains largely unexplored how various factors (design, habitat quality, landscape context) interact to determine the occurrence of butterflies in flower strips. Therefore, butterflies were surveyed in 15 flower strips differing in age (first and second growing season). Flower strips were compared with 15 field margins, which were adjacent to arable land and were dominated by grasses. The field studies were conducted during two summers (2013, 2014) in Lower Saxony (Germany). Additionally, based on a literature study, 17 environmental variables likely to be decisive for the occurrence of butterflies were identified and recorded during these field studies or analyzed in GIS. Supported by a PCA, 8 environmental variables for flower strips and 7 for field margins, were selected and included in linear mixed-effects models in order to calculate their effect on butterflies.

We documented 19 butterfly species and 1,394 individuals in the flower strips and 13 species and 401 individuals in the field margins. The number of flowering plant species was the key factor for the occurrence of butterflies - both in flower strips and field margins. The diversity of the surrounding landscape (Shannon-Index H) had an additional significant influence on butterflies in flower strips, with more species and individuals being observed on areas with a lower Shannon-Index.

Number of flowering plant species is the key driver of butterfly diversity and abundance, which improves the habitat quality of flower strips in agricultural landscapes. In order to promote butterflies optimally, flower strips must have a good supply of flowers even over several years. This requires careful design and management, as flower supply often decreases with increasing age of the flower strips. The study indicates that flower strips have a particularly high effect in structurally simple landscapes.

1. Introduction

Recent declines of pollinators are startlingly evident (Potts et al., 2010). Since butterflies react rapidly to changes in habitat quality, they are especially affected (Hambler et al., 2011; Thomas et al., 2004) and the decline has already been observed in previously widespread butterfly species (Fox et al., 2015). Land use intensification is a main driver of the decline of butterfly diversity (Brittain et al., 2010; Fox et al., 2015; Warren et al., 2001).

In principle, flower strips, defined here as strips on arable land which are managed specifically by sowing a species-rich seed mixture (e.g. Supplementary data: Table A.1) with the aim of creating flower-rich and structurally rich habitats, can counteract biodiversity loss in temperate farmland (Haaland et al., 2011; Ouvrard et al., 2018; Sutter et al., 2018; Uyttenbroeck et al., 2015). Studies show that flower strips can especially promote butterflies in the agricultural landscape: Flower strips were more species-rich and/or more individual-rich than other habitats of the agricultural landscape (Aviron et al., 2011: flower strips vs. conventional fields, Haaland and Bersier, 2011: flower strips vs. extensively used meadows, Haaland and Gyllin, 2010: flower strips vs. greenways). For that, however, it is important that certain basic conditions and design options are taken into account, such as the composition of seed mixture, suitable soil conditions, the management of the flower strips and the life span of the flower strips (Aviron et al., 2011; Haaland and Bersier, 2011;

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Haaland and Gyllin, 2010).

There is still a considerable need for research on optimizing the design of flower strips (Holland et al., 2015; Uyttenbroeck et al., 2017) and it is still unclear which factors determinate species richness and abundance of butterflies in flower strips. Previous studies have revealed that various factors at the local level influence the occurrence of butterflies in flower strips (e.g. Woodcock et al., 2014: flora and management; Dollar et al., 2013: management; Pywell et al., 2011a: flora, management and age). For other habitats, however, several studies have already shown that both local scale habitat characteristics, as well as land use and connectivity with surrounding landscapes, are relevant factors for the occurrence of butterflies (Luppi et al., 2018; Sybertz et al., 2017; Delattre et al., 2010; Ouin and Burel, 2002). However, the transfer of results from other semi-natural habitats to flower strips is inadequate, as flower strips represent new landscape elements (Wagner et al., 2014). In Germany and Switzerland, the seed mixtures of flower strips generally do not contain any grass species (Haaland et al., 2011). Hence, they can hardly be equated with species-rich meadows or field margins, defined here as occasionally mown, grass-dominated strips adjacent to arable land. Flower strip species composition also differs from perennial tall herb communities and ruderal. Therefore, the explicit analysis of flower strips is necessary, especially when examining how a variety of different factors of different scales (design of flower strips, habitat quality and landscape context) interact to determine the occurrence of butterflies.

In light of these issues, the aim of our study was to find and to compare key factors that determine butterfly occurrence in flower strips and field margins. Utilizing linear mixed-effect models we wanted to answer the following research questions: (1) Which factors (site-specific characteristics, including growing season or vegetation, as well as heterogeneity and connectivity of the surrounding landscape) determine the occurrence of butterflies (species richness and abundance) in flower strips? (2) Which of these factors determine the occurrence of butterflies in field margins? (3) How do flower strips affect adjacent areas? (4) How can this knowledge be used to promote the occurrence of butterflies in the agricultural landscape?

2. Materials and methods

2.1. Study sites

The study was carried out in the vicinity of Zeven (district of Rotenburg, Lower Saxony, Germany $(53^{\circ}17'38.47 \text{ N}, 9^{\circ}16'33.51 \text{ E})$, Fig. 1). About 70% of the districts area is used for agriculture (LSN, 2018).

Ten flower strips, with five having repeat measurements, resulting in a total sample size of 15 flower strips, and 15 field margins were examined (Table 1). The flower strips studied were created on maize fields by sowing a species-rich seed mixture of 13 perennial flowering plant species (Supplementary data: Table A.1). During their life span of 1.5 years, no tillage, fertilization, or plant protection measures were permitted on the flower strips. Field margins were located next to maize fields, were grass-dominated and were mown occasionally. Concerning the flower strips, we focused on different ages (growing seasons, Table 1). The flower strips of the first growing season (2013) were also examined in 2014 when they were in the second growing season. Given the high degree of interannual variability in butterfly occurrences, five new flower strips in their first growing season were additionally investigated in 2014. All flower strips were 6 m wide (Table 1), were created by the same 'Rotenburger seed mixture 2013' (Supplementary data: Table A.1) and were located between maize fields and unsealed farm tracks. The field margins were examined in 2014 and were between 1 and 4 m wide

Table 1

Overview of the investigated flower strips and field margins.

| a) Flower strips | | | | | |
|--|------------------------------------|--|---|--|--|
| Year | Width | Growing season | Number of study sites | | |
| 2013 | 6 m | 1. Growing season | n = 5 | | |
| 2014 | 6 m | 1. Growing season | n = 5 | | |
| 2014 | 6 m | 2. Growing season | n = 5 | | |
| | | | | | |
| b) Field r | nargins | | | | |
| b) Field r Year | nargins Width | Adjacent area | Number of study sites | | |
| b) Field r Year 2014 | nargins Width 2–4 m | Adjacent area Maize fields | Number of study sites $n = 5$ | | |
| b) Field r Year 2014 2014 | nargins Width 2–4 m 1–2 m | Adjacent area Maize fields Flower strips 1. growing season | Number of study sites n = 5 n = 5 | | |

repetition measurements of the 5 flower strips from 2013.



Fig. 1. The study area district Rotenburg (Wümme) (grey section in the small picture) is located in Lower Saxony (outlined in black), Germany. The study sites are located in the vicinity of Zeven (data basis: GeoBasis-DE / BKG, 2017; MU Nds., 2018).

(Table 1). The field margins were also located next to maize fields. Moreover, we additionally investigated field margins next to flower strips of the first and second growing season in order to analyze the influence of neighboring flower strips.

2.2. Response variables: number of butterfly species and individuals

The butterflies were surveyed between June and August 2013 and 2014 by standardized visual observations of imagines using line-transects following Settele (1999). A transect of 125 m was sampled for each study site. The width of the transect was based on the width of flower strips of 6 m. Due to the narrower field margins, the transect width was based on the width of the respective field margin (1–4 m, Table 1). Consecutive dates of surveys were blocked in rounds (2–7 days in each round). In 2013, we conducted 4 rounds with a total of 8 inspections per study site. In 2014, there were 5 rounds with a total of 11 inspections per study site.

If individual species could not be determined, they were included in the protocols as species complexes (*Artogeia rapae/napi*, *Thymelicus lineola/sylvestris*) or families (Blues). For taxonomic determination we used Settele and Steiner (2009) and Tolman and Lewington (2009). The nomenclature follows the latter.

The number of recorded butterfly species (as a proxy for species richness, defined by the total sum of recorded species) and the number of individuals (as a proxy for abundance, defined by the total sum of recorded individuals) were used as response variables.

2.3. Explanatory variables: characteristics of the habitat type and the surrounding landscape

2.3.1. Determination, description and recording of variables affecting butterflies

A systematic literature review within the Web of Science database

was carried out using the search terms 'butterfl*' and 'flower strip*' or 'field margin' in the categories 'title', 'summary' and 'keywords' for the publication years from 2014 to 2018 (*stands for any group of characters). Further publications were supplemented using the snowball system, whereby additional references were included from the bibliographies of the Web of Science articles. Based on the literature review, we identified the main parameters that are demonstrably influential for the occurrence of butterflies in flower strips, field margins or comparable structures in the agricultural landscape (Table 2). As the spatial scale plays an important role in the analysis and evaluation of biodiversity (Ekroos et al., 2013; Ekroos and Kuussaari, 2012; Gabriel et al., 2010), we structured the variables into two scales.

Based on the results of the literature review (Table 2), we selected seven main parameters of different spatial scales which could be described by different variables (Table 3). In total, we recorded 16 variables in flower strips and 11 variables in field margins.

To record the vegetation structure (Table 3), five sample plots (1 \times 1m) were randomly distributed on each study site (random quadrats, Traxler, 1997) and surveyed in both years, four times between June and August. At each collection date, the location of sample plots was re-measured (temporary plots, Traxler, 1997). The floristic features (Table 3: Abundant flowering of the study sites) were recorded at the same time as the vegetation structure and also with the same number of repetitions. The variables of landscape heterogeneity and of connectivity were calculated by GIS-analyses within a 1 km buffer around each study site (Table 3). The landscape heterogeneity (Shannon Diversity Index, Shannon Evenness Index and Edge Density; Lang and Blaschke, 2007) are based on CORINE Land Cover Data which were generalized to a minimum size of 10 hectares (GeoBasis-DE / BKG, 2012). The connectivity (effective mesh size; Jaeger, 2000) was estimated from digital topographic maps and orthophotos (MU Nds., 2018). On this basis, all linear elements representing potential habitats or corridors for butterflies

Table 2

Variables determining butterfly occurrences in flower strips, field margins or comparable structures in the agricultural landscape. References indicating investigated flower strips are underlined.

| Main parameters (group) | Sources |
|---|---|
| Year of investigation | Aviron et al. (2011); Aviron et al. (2007a); Feber et al. (1996) |
| Scale: Study site | |
| Habitat type | Aviron et al. (2011); Aviron et al. (2007a); Berg et al. (2011); Feber et al. (1996); Haaland and Bersier |
| | (2011); Kuussaari et al. (2007); Weibull et al. (2003); Wix and Reich (2018) |
| Habitat quality | Dover and Settele (2009); Ekroos and Kuussaari (2012); Kuussaari et al. (2007) |
| Patch geometry (e.g. width, length, shape, size) | Clausen et al. (2001); Cole et al., (2015); Delattre et al. (2010); Dover (1996); Dover and Settele (2009); |
| | Field et al. (2007), 2006, 2005; Korpela et al. (2013); Kuussaari et al. (2007); Pywell et al. (2004); |
| | Saarinen et al. (2005); Skórka et al. (2013); Sparks and Parish (1995); Sybertz et al. (2017); Wix and |
| | Reich (2018) |
| Insolation | Clausen et al. (2001), Dover (1996), Pywell et al. (2004) |
| Growing season/Age | Haaland and Bersier (2011); Korpela et al. (2013); Pywell et al. (2011a); Wix and Reich (2018) |
| Vegetation structure (e.g. amount of trees and shrubs, vegetation high) | Berg et al. (2013): Clausen et al. (2001): Dover et al. (2000): Haaland and Gyllin (2010): Pywell et al. |
| | (2004); Sparks and Parish (1995); Sybertz et al. (2017); Wix and Reich (2018) |
| Flora (e.g. floral composition, grass-herb-ratio, abundant flowering, | Berg et al. (2013); Clausen et al. (2001); Cole et al., (2015); Dover et al. (2000); Dover (1996); Ekroos |
| nectar plants and larval food plants) | et al. (2008); Feber et al. (1996); Field et al. (2006); Gabriel et al. (2010); Haaland and Bersier (2011); |
| | Haaland and Gyllin (2010); Kuussaari et al. (2007); Lebeau et al., (2016); Meek et al. (2002); Noordijk |
| | et al. (2009): Pywell et al. (2011a): Pywell et al. (2004): Saarinen et al. (2005): Saarinen (2002): Skórka |
| | et al. (2013): Sparks and Parish (1995): Sybertz et al. (2017): Woodcock et al., (2014): Wagner et al. |
| | (2014): Wix and Reich (2018) |
| Management (e.g. grazing, (partial) mowing (time), removal of cuttings, | Aviron et al. (2007b): Dollar et al. (2013): Feber et al. (1996): Giuliano et al., (2018): Kruse et al., |
| use of pesticides) | (2016): Noordijk et al. (2009): Pywell et al. (2011a): Saarinen (2002): Snoo (1999): Snoo et al. (1998): |
| I | Valtonen et al. (2006): Woodcock et al., (2014) |
| Scale: Surrounding landscape | |
| Land use of the adjacent field (e.g. habitat type, type of crop) | Dover (1996); Pywell et al. (2004); Saarinen et al. (2005); Sybertz et al. (2017) |
| Management of the adjacent field (e.g. organic farming) | Clausen et al. (2001); Jonason et al. (2011); Rundlöf et al. (2008); Rundlöf and Smith (2006); Sybertz |
| | et al. (2017); Taylor and Morecroft (2009) |
| Shelter by adjacent structures (e.g. hedges, bushes, buffer strips) | Clausen et al. (2001); Dover et al. (1997); Dover (1996); Pywell et al. (2004); Sybertz et al. (2017) |
| Landscape heterogeneity, landscape type, landscape composition | Aviron et al. (2007a), 2011; Berg et al. (2011); Dainese et al., (2015); Dover and Settele (2009); Ekroos |
| (percentage of organically managed farmland in a district) | and Kuussaari (2012); Gabriel et al. (2010); Haaland and Bersier (2011); Jonason et al. (2011); Korpela |
| | et al. (2013); Ouin and Burel (2002); Rundlöf et al. (2008); Rundlöf and Smith (2006); Šálek et al., |
| | (2018); Skórka et al. (2013); Sybertz et al. (2017); Toivonen et al. (2017), 2016, 2015; Weibull et al. |
| | (2003); Weibull et al. (2000) |
| Connectivity (e.g. isolation, fragmentation, barriers) | Aviron et al. (2007b); Brückmann et al. (2010); Delattre et al. (2013), 2010; Dover and Settele (2009); |
| | Ouin and Burel (2002) |

Table 3

Overview of all explanatory variables recorded for each flower strip or field margin. Non-bold x in columns FS (Flower strips) and FM (Field margins) indicate variables that were recorded in the respective habitat type. Bold variables and X in columns FS (Flower strips) and FM (Field margins) are the selected variables (for selection see chapter 2.4).

| Main parameter | Variable | Abbreviation | Description | FS | FM |
|--|--|---------------------|--|----|----|
| Characteristics of the survey | Year | year | Year in which the surveys were carried out. | X | |
| Scale: Study sites General characteristics of the study site | Width | width | Absolute width of the study site [m]. | x | x |
| | Age/Growing season | gro_sea | The life span of flower strips. Variable coded in R: $1 =$ flower strips which are in their first growing season, $2 =$ flower strips which are in their second | x | |
| | Adjacent area | adj_area | growing season Habitat type adjacent to the study site. Variable coded in R: $0 =$ maize field, 1 = flower strip which are in their first growing season, $2 =$ flower strip which are in their second growing season | | x |
| Vegetation structure of the study sites | Open-ground proportion | op_ground | Average open-ground proportion [%] of all five sample plots $(1 \times 1m)$ of a study site. This variable was recorded at each round. | X | x |
| | Dominant height of vegetation | dom_veghigh | For each vegetation layer, the average of the vegetation cover and the vegetation height was formed from the five sample squares. The average vegetation height of the layer with the highest coverage represents the dominant vegetation height in cm. This variable was recorded at each round | x | X |
| | Maximum height of vegetation | max_veghigh | Maximum value of the vegetation height [cm] of all five sample plots $(1 \times 1 m)$ of a study site. This variable was recorded at each round. | X | x |
| Abundant flowering of the study sites | Number of flowering species | no_flower | Sum of all recorded flowering, herbaceous plant species (the plant species of the seed mixture included) with a medium abundance on the study site (from | X | x |
| | Cover level/Stand cover of flowering species | cover_flower | a cover level of 10%). This variable was recorded at each round. Cover level [%] of all recorded flowering, herbaceous plant species (the plant species of the seed mixture included). This variable was recorded at each round. | x | x |
| Growth of the plants from the seed mixture | Total number of plant species | no_seedmix_total | Sum of recorded plant species of the seed mixture (from a cover level of 1%). This variable was recorded at each round. | x | |
| | Number of flowering plant species of the seed mixture | no_seedmix_flow | Sum of recorded flowering plant species of the seed mixture. This variable was recorded at each round. | х | |
| | Cover level/Stand cover of plant species of the seed mixture | cover_seedmix_total | Sum of cover level [%] of all recorded plant species of the seed mixture (from a cover level of 1%). This variable was recorded at each round. | x | |
| | Cover level/Stand cover of flowering plant species of the seed mixture | cover_seedmix_flow | Sum of cover level [%] of all recorded flowering plant species of the seed mixture. This variable was recorded at each round. | x | |
| Scale: Surrounding landscape | Diversity of habitat types | chan H | Shannon Diversity Index (H) calculated from babitat types within a 1 km | v | v |
| Landscape necerogenenty | Diversity of habitat types | shan_n | buffer around a study site. $H = \sum_{i=1}^{m} pi^{i*} npi$ | л | А |
| | | | P = proportion of the habitat type i, m = number of habitat types (Lang and Blaschke, 2007). The proportion of each occurring habitat type within a 1 km buffer around a study site was determined on the basis of Corine Land Cover 10.000 million of the basis of Corine Land Cover 10.000 million of the basis of Corine Land Cover 10.000 million of the basis of Cover Land Cover 10.000 million of the basis of Cover Land Cover 10.000 million of the basis of Cover 10.0000 million of the basis of Cover 10.00000 million of the basis of Cover 10.00000 million of the basis of Cover 10.00000 million of the basis of Cover 10.00000000000000000000000000000000000 | | |
| | Evenness of habitat types | shan_E | 10 (GeoBasis-DE / BKG, 2012) Shannon Evenness Index (Even) calculated from habitat types within a 1 km buffer around a study site. | x | x |
| | | | $Even = \frac{H}{\ln m}$ H = Shannon Diversity Index (see description of shan_H), m = number of habitat types (Lang and Blaschke, 2007). The proportion of each occurring habitat type within a 1 km buffer around a study site was determined on the basis of Corine Land Cover 10 (GeoBasis-DE / BKG, 2012) | | |
| | Structural diversity | ed | Edge Density: Total edge within a 1 km buffer around a study site in meters per hectare (Lang and Blaschke, 2007). The edge density was determined on the basis of Corine Land Cover 10 (GeoBasis-DE / BKG, 2012) and calculated with the Extension for ArcGIS "V I ATE 2.0 heta" | x | x |
| Connectivity | Effective mesh size | mesh | Effective mesh size calculated from potential habitats and corridors for butterflies within a 1 km buffer around a study site in ha. | x | x |
| | | | $mesh = \frac{1}{At} \sum_{i=1}^{n} A_i^2$ | | |
| | | | $n =$ number of patches, $A_i =$ sizes of the n patches; $A_t =$ total area of the 1 km buffer (Jaeger, 2000). The effective mesh size was determined on the basis of digital topographic maps, orthophotos and Corine Land Cover 10 (GeoBasis-DE / BKG, 2012; MU Nds., 2018) | | |

2.4. Selection of explanatory variables

outside villages were digitized (e.g. country lanes, field margins or strips along ditches). This dataset was intersected with potentially suitable butterfly habitats from the CORINE Land Cover Dataset (relevant for the study area: pasture, meadows and other permanent grasslands under agricultural use (code 231), natural grassland (code 321), moors and heathland (code 322), GeoBasis-DE / BKG, 2012).

To reduce the number of variables according to the sample size and to avoid redundancy, a principal component analysis (PCA) after standardization (Quinn and Keough, 2014) was carried out for each of the main parameters (Figs. 2 and 3). In order to clearly represent the ecological relationships of individual variables and to facilitate the practical application of the model (survey of the variables in the field or



Fig. 2. Principal component analysis for potential fixed-effects of the three main parameters (vegetation structure, abundant flowering and surrounding landscape) for the model of flower strips (numbers indicate single study sites on individual investigations). For abbreviations of the variables see Table 3.



Fig. 3. Principal component analysis for potential fixed-effects of the three main parameters (vegetation structure, abundant flowering and surrounding landscape) for the model of field margins (numbers indicate single study sites on individual investigations). For abbreviations of the variables see Table 3.

implementation of the variables in practice), we chose individual, non-redundant variables of the different grouped variables and did not use the components of the PCA. Variables associated with the first and second component were selected (Supplementary data: Table B.1-B.6, Figs. 2 and 3). When groups of several variables showed association with one component (e.g. Fig. 2, Abundant flowering), the variable selected from each group was the one which could be surveyed in the field with least effort (e.g. number of flowering species rather than cover of flowering species). Therefore, we were able to exclude seven highly redundant variables for flower strips and four for field margins.

As the annual fluctuation of butterflies is crucial and as the life span (age) of flower strips can be controlled during the creation of flower strips, their influence was assessed as decisive. With regard to the results of the literature search (Table 2), the different widths of the field margins (Table 1) were also classified as relevant variables. The adjacent area of the field margins (maize or flower strip, Table 1) was crucial for evaluating the impact of flower strips on these areas. Therefore, once their influence was proved significant, the four variables (year and growing season for flower strips, width and adjacent area for field margins) were consistently included as fixed effects in the respective models.

Finally, eight variables were selected for the models of flower strips and seven variables for the models of field margins (written in bold and marked with '**X**' in Table 3 in columns FS or FM).

2.5. Random-effects

Some flower strips were examined in 2013 as well as in 2014 and they cannot be considered as independent variables without further criteria. Since the year of the investigation (fluctuation, weather conditions) has a decisive influence on the occurrence of butterflies (Table 2), it was used as a variable to define independent data. Furthermore, as the butterflies were recorded on blocked dates and against the background of phenological changes, the consecutive dates (defined as rounds) were taken into account in the statistical analyses. Therefore, we considered one study site for each round of the respective year of investigation as independent data basis for the models (Table 4). We thus included as random effects: the variance between study sites, the interaction of study site and year as well as the interaction of study site, year and round: (1|site) + (1|site:year) + (1|site:year:rd). Since the field margins were only examined in 2014, the year variable could be removed from the data structure of the random-effects in the models for field margins.

2.6. Linear mixed-effects models

We used linear mixed-effects models fit by maximum likelihood (ML) estimations. The exploratory graphics generated showed a right-skewed distribution. For the assumption of normally distributed data (residuals and random-effects), the number of species and individuals was log-transformed (log (y+1)). Time-correlated measurements for the time series of the individual study sites were assumed (corAR1-structure).

The selection of the fixed effects was based on forward hierarchical selection (Kuckartz et al., 2013; Quinn and Keough, 2014), involving the sequential addition of the effects in a pre-defined way. The order was directed from local scale to that of landscape scale. Based on an ecological background, the main parameters that were expected to have the greatest influence on the butterflies were included first. By the order of inclusion, the additional effect of a variable could be tested. For model comparison and selection, likelihood ratio tests (LRT, Bolker et al., 2009; Zuur et al., 2009) and corrected Akaike information criterion (AICc) were used (Burnham and Anderson, 2002). 95% confidence intervals were computed for the fixed effect parameters of the final mixed effect models

| Table 4 | | | |
|----------|--------|-------------|-------|
| Overview | of the | random-effe | ects. |

| Variable | Abbreviation | Description |
|----------|--------------|---|
| Year | year | Year in which the surveys were carried out |
| Round | rd | The butterflies were recorded on blocked dates. The |
| | | blocked appointments were summarized in time as |
| | | rounds of 2–7 days |
| Study | site | Name of the study site |
| site | | |

fitted by restricted maximum likelihood (REML). For these final models, the assumptions (homoscedasticity, normality of residuals and random-effects) were visually inspected by residual plots and Q-Q plots. As compared to the untransformed data, the log-transformed data showed no or reduced deviations from normal distribution (QQ-normal-plot), and residual vs. fitted plots showed no heteroscedasticity.

Analyses were calculated in the R language and environment (RStudio Team, 2016). For linear mixed-effect models the package 'nlme' was used (Pinheiro et al., 2018). Post-hoc comparisons of means between different types of adjacent areas at pre-specified coavariate values (width) were performed using R package 'emmeans' (Lenth, 2018). AICc were calculated with the package 'MuMIn' (Kamil, 2018). Some graphics were created with 'ggplot2' (Wickham, 2016).

3. Results

3.1. Butterflies in flower strips and field margins

In total, we recorded 19 butterfly species and 1,795 individuals (Supplementary data: Table C.1-C.2). Only in the flower strips could all 19 butterfly species be detected. Seven species were solely documented here: *Agrodiaetus amandus, Colias crocea, Lycaena phlaeas, Pieris brassicae, Polygonum c-album, Polyommatus icarus, Vanessa cardui*). With 1,394 individuals compared to 401 individuals, three times as many individuals were observed in the flower strips as in the field margins. No endangered species were found (Reinhardt and Bolz, 2011) and only one grassland specialist according to Swaay et al. (2006), *Agrodiaetus amandus*, could be recorded. Considerable proportions of the recorded butterflies were classified as generalists at a European level (14 species) or as using a variety of biotopes (4 species).

3.2. Effects of habitat quality and surrounding landscape on butterflies

3.2.1. Interactions between selected explanatory variables

In general, we recorded a higher abundance of flowering species in the flower strips in the first growing season than in the second growing season. Though the medians were similar (first growing season: 4 species, second growing season: 3 species), the maximal number of species differed with 11 species recorded in flower strips in the first growing season compared to only 5 species in flower strips in the second growing season.

With an average width of 3 m (median), the field margins next to maize fields were much wider than those next to flower strips (median of 1.5 m; both, flower strips in the first and second growing season).

3.2.2. Impact of fixed-effects on butterflies in flower strips

Sequential testing of a total of eight explanatory variables in different model comparisons (Table 5) identified three variables which form wellsupported predictors of the occurrence of butterflies (species richness and abundance) in flower strips: the growing season, the number of flowering species and the diversity of habitat types.

The model including the variables 'growing season' and 'year of investigation' showed significant influence on both the number of species and the number of individuals in contrast to the model including only the 'year of investigation' (Table 5, Test mod1s/mod1i vs. mod2s/mod2i). In both years the numbers of species and individuals were similar (median: 2 species (2013 and 2014), 7 individuals (2013) and 7.5 individuals (2014), range: 0–6 species (2013) and 0–8 species (2014), 0–35 individuals (2013) and 0–42 individuals (2014)). On account of this fact we tested the model without the influence of the year (mod3s/i): Excluding the 'year of investigation' showed no significant impact (Test mod2s/mod2i vs. mod3s/mod3i). For our investigation period it can

Table 5

Sequentially fitted linear mixed-effects models to analyze the occurrence of butterfly species (a) and individuals (b) in flower strips and the sequential likelihood ratio tests of added fixed effects (LRT: Test statistic of likelihood ratio test) with indication of significance (p-value, sign). Degree of freedom (df) and corrected Akaike information criterion (AICc) are provided for each fitted model. The selected variables/models used in the subsequent model comparisons (LRT, Test), as well as the models with lowest AICc (for the respective analysis step) are written in bold letters. The final model selected for use is underlined.

Flower strips a) Response variable: Total number of butterfly species (log (1 + species)) Model Variables Test LRT p-value sign. df AICc Study design mod1s 7 236.4 ~ vear ** mod2s mod1s vs. mod2s 7.0198 0.0081 8 231.6 ~ year + gro_sea mod3s ~ gro_sea mod2s vs. mod3s 2.0624 0.1510 n.s. 7 231.4 Abundant flowering *** mod4s gro_sea + no_flower mod3s vs. mod4s 30.9056 <.0001 8 202.7 mod5s ~ gro_seaf + no_flower + no_seedmix_flow mod4s vs. mod5s 1.4260 0.2324 9 203.6 n.s. Vegetation structure 2.5624 0.1094 202.4 mod6s \sim gro sea + no flower + op ground mod4s vs. mod6s n.s. 9 mod7s \sim gro_sea + no_flower + max_veghigh mod4s vs. mod7s 2.4305 0.1190 9 202.6 n.s. Surrounding landscape mod8s ~ gro_sea + no_flower + shan_H mod4s vs. mod8s 5.4554 0.0195 9 199.5 \sim gro_sea + no_flower + shan_H + mesh 10 mod9s mod8s vs. mod9s 0.8519 0.3560 n.s. 201.0 b) Response variable: Total number of butterfly individuals (log(1+indiv)) Model Variables Test LRT p-value sign. df AICc Study design mod1i ~ year 7 405.4 mod2i mod1i vs. mod2i 8.1659 0.0043 ** 8 399.5 \sim year + gro sea 0.3861 mod3i mod2i vs. mod3i 0.7511 398.0 ~ gro_sea 7 Abundant flowering 26.8662 <.0001 *** 8 373.4 mod4i r gro_sea + no_flower mod3i vs. mod4i mod5i gro_sea + no_flower + no_seedmix_flow mod4i vs. mod5i 1.1671 0.2800 9 374.5 n.s. Vegetation structure mod6i ~ gro_sea + no_flower + op_ground mod4i vs mod6i 2 5665 0 1092 n s 9 373 1 mod7i mod4i vs. mod7i 0.2113 9 374.1 \sim gro sea + no flower + max veghigh 1.5627 n.s. Surrounding landscape ** mod8i \sim gro sea + no flower + shan H mod4i vs. mod8i 7.8347 0.0051 9 367.8 mod8i vs. mod9i 1.1399 0.2857 10 mod9i ~ gro_sea + no_flower + shan_H + mesh n.s. 369.0 Signif. codes: ***p < 0.001; **p < 0.01; *p < 0.05; $\cdot p < 0.10$; n.s. $p \ge 0.10$.

therefore be concluded that the 'growing season' has a more decisive influence on butterfly occurrence than the 'year'. Since the model including only the 'growing season' (mod3s/mod3i) had a lower degree of freedom and a lower AICc-value than the model including 'growing season' and 'year' (mod2s/mod2i), the model mod3s/mod3i was selected for the following model comparisons. Adding the 'number of flowering species' to the model showed a highly significant difference on both, number of butterfly species and individuals (mod4s, mod4i, p < p0.0001). In addition, this factor considerably improved the model quality in comparison to the previous model (AICc: 231.4 (mod3s) to 202.7 (mod4s); 398.0 (mod3i) to 373.4 (mod4i)). Neither the 'number of flowering species of the seed mixture' nor the vegetation structure ('open-ground proportion' and 'maximum high of vegetation') had a significant additional influence on the occurrence of butterflies (mod5smod7s; mod5i-mod7i). With regard to the surrounding landscape, including the 'diversity of habitat types' in the models (mod8s, mod8i) leads to a significant difference and to an improvement of the model quality compared to the previous models (mod4s, mod4i). The 'effective mesh size' had no significant influence neither on the number of species nor on the number of individuals (mod9s, mod9i).

The median number of species and individuals in flower strips in the first growing season (3 species, 8 individuals) was higher than the one in

the flower strips in the second growing season (1 species, 3 individuals, Fig. 4). The number of butterfly species of flower strips in the first growing season decreased by a factor of 0.65 or increased by a factor of 1.20 (95% CI) in comparison to the flower strips in the second growing season. In terms of the number of individuals, the factor was 0.39–1.22 (95% CI).

Greater flowering resources increased species richness as well as abundances (Fig. 4). When the number of flowering species increased by 1, the number of butterfly species increased on average by 10-19% and the number of individuals by 17-36% (95% CI).

Here, the interaction between growing season and number of flowering resources has to be taken into account (Chapter 3.2.1). We recorded more butterflies (species and individuals) in flower strips with a high abundance of flowering species (Fig. 5). Furthermore, only the flower strips in the first growing season showed a high number of flowering resources with over 6 plant species. Overall, the flower strips in the first growing season were more species- and individual-rich than those in the second growing season in terms of butterflies and flowering resources (the latter only in the number of species).

The landscape heterogeneity had a negative, significant effect on butterfly occurrences. The Shannon-Index H could not be as clearly visualized with the raw data (Fig. 4), as this effect was estimated as an



Fig. 4. Boxplots and scatterplots of the fixed-effects with significant influence of the number of butterfly species and individuals (Table 5) in flower strips (n = 150 observations: 8 inspections per study site in 2013, 5 study sites in 2013 and 11 inspections in 2014 per study site, 10 study sites in 2014, see Table 1). In order to make the graphic clearly legible, the dots in the scatterplots were jittered.



Fig. 5. Relation between the number of species or individuals in flower strips and the number of flowering species subdivided by the growing season of the flower strips (open circles = first growing season, closed circles = second growing season; n = 150 observations: 8 inspections per study site in 2013, 5 study sites in 2013 and 11 inspections in 2014 per study site, 10 study sites in 2014, see Table 1). In order to make the graphic clearly legible, the dots were jittered.

additional influence, i.e. after the influence of the effects previously included in the model - growing season, number of flowering plants were calculated out of the data. The scatterplots indicate that high numbers of butterfly species and individuals accumulate at low Shannon-Index H values. However, some high species and individual numbers also occur at a high Shannon-Index H. Finally, if the Shannon-Index H increased by 1 unit, butterfly species decreased by 47% or remained almost the same with only a reduction of 1% (95% CI). Concerning the number of individuals, the negative effect was more evident by a factor decrease of 0.28–0.88 (95% CI). Because the Shannon-Index H itself encompassed only a very limited range (0.42–1.61), the increase in the slope of this variable by 1 was a substantial rise and so the CI varied in such a high range.

3.2.3. Impact of fixed-effects on butterflies in field margins

Sequential testing of a total of seven explanatory variables in different model comparisons (Table 6) identified three variables that are strongly associated with the occurrence of butterflies (species richness and abundance) in field margins: width, adjacent area and abundant flowering.

The field margins along the flower strips were significantly narrower than the field margins along maize fields (Chapter 3.2.1), requiring an examination of the interaction between 'width' and 'adjacent area'. The interaction between 'width' and 'adjacent area' described the occurrence of butterflies (number of species and individuals) better than their additive inclusion (modAs/modAi vs modBs/modBi, Table 6). The inclusion of the 'number of flowering species' in these models with the interaction (modCs, modCi) showed a highly significant influence (p < 0.0001) and distinctly improved the model quality (the number of species and individuals). None of the further factors concerning the vegetation structure or the surrounding landscape indicated a significant difference or improvement of model quality compared to the model modCs or modCi.

The maximum number of species and in particular the number of

individuals in field margins along flower strips (independent of the growing season) were higher than these in field margins along maize fields (Fig. 6). In comparison to field margins along flower strips in the first growing season, the number of observed species in field margins along maize fields differed by a factor of 0.72–1.97, the number of individuals by a factor of 0.57–2.69 (95% CI). For the comparison between the field margins along maize fields and the field margins along flower strips in the second growing season, the factor was 0.86–2.88 (number of species) or rather 0.70–4.56 (number of individuals). Furthermore, the field margins along flower strips in the first growing season were more species-rich and individual-rich than those along flower strips in the second growing season. These two types differed by a factor of 0.78–2.24 (species) or 0.64–3.28 (individuals).

The field margins next to the maize fields (width of 2-4 m) were much wider than the field margins along the flower strips (1–2 m, Fig. 7). Nevertheless, the highest numbers of butterflies (especially number of individuals) were recorded in the much narrower field margins along the flower strips.

The supply of flowering resources is strongly associated with butterfly abundance: A higher number of flowering plants attracted more butterfly species and individuals (Fig. 6). In average, with one additional flowering species the number of butterfly species increased by 14–44% and the number of individuals by 28–84% (95% CI).

4. Discussion

4.1. Species range of butterflies in flower strips and field margins

In the flower strips as well as in the field margins we have mainly recorded generalists. Other studies also indicated that flower strips usually promote generalists (e.g. Haaland et al., 2011). However, although no endangered species and only one grassland specialist could be detected in this study, this is not due to a minor effect of the flower strips or field margins. Rather, it has to be considered that the regional

Table 6

Sequentially fitted linear mixed-effects models to analyze the occurrence of butterfly species (a) and individuals (b) in field margins and sequential likelihood ratio tests of added fixed effects (LRT: Test statistic of likelihood ratio test) with indication of significance (p-value, sign). Degree of freedom (df) and corrected Akaike information criterion (AICc) are provided for each fitted model. The selected variables/models used in the subsequent model comparisons (LRT, Test), as well as the models with lowest AICc (for the respective analysis step) are written in bold letters. The final model selected for use is underlined.

| a) Response varia | able: Total number of butterfly species (log $(1 + \text{species}))$ | | | | | | |
|--|---|---|---|---|---|---------------------------------------|--|
| Model | Variables | Test | LRT | p-value | sign. | Df | AICc |
| Study design | | | | | | | |
| modAs | \sim width + adj_area | | | | | 10 | 255.6 |
| modBs | ~ width * adj_area | modAs vs. modBs | 7.6489 | 0.0218 | * | 8 | 252.5 |
| Abundant flowerir | Abundant flowering | | | | | | |
| modCs | ~ width * adj_area + no_flower | modBs vs. modCs | 17.2564 | <.0001 | *** | 11 | 237.5 |
| Vegetation structu | re | | | | | | |
| modDs | \sim width * adj_area + no_flower + op_ground | modCs vs. modDs | 0.0003 | 0.9867 | n.s. | 12 | 239.8 |
| modEs | \sim width * adj_area + no_flower + dom_veghigh | modCs vs. modEs | 0.0502 | 0.8228 | n.s. | 12 | 239.8 |
| Surrounding lands | scape | | | | | | |
| modFs | \sim width * adj_area + no_flower + shan_H | modCs vs. modFs | 0.0355 | 0.8506 | n.s. | 12 | 239.8 |
| modGs | \sim width * adj_area + no_flower + mesh | modCs vs. modGs | 0.5813 | 0.4458 | n.s. | 12 | 239.3 |
| b) Response variable: Total number of butterfly individuals (log(1+indiv)) | | | | | | | |
| b) Response vari | able: Total number of butterfly individuals (log(1+indiv)) | | | | | | |
| b) Response vari Model | able: Total number of butterfly individuals (log(1+indiv)) Variables | Test | LRT | p-value | sign. | df | AICc |
| b) Response varia Model Study design | able: Total number of butterfly individuals (log(1+indiv)) Variables | Test | LRT | p-value | sign. | df | AICc |
| b) Response varia Model Study design modAi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area | Test | LRT | p-value | sign. | df 8 | AICc 370.8 |
| b) Response vari Model Study design modAi modBi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area | Test modAi vs. modBi | LRT 6.10398 | p-value 0.0473 | sign. | df 8 10 | AICc 370.8 369.2 |
| b) Response vari Model Study design modAi modBi Abundant flowerin | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area | Test modAi vs. modBi | LRT 6.10398 | p-value 0.0473 | sign. * | df 8 10 | AICc 370.8 369.2 |
| b) Response varia Model Study design modAi modBi Abundant flowerin modCi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area ng ~ width * adj_area + no_flower | Test modAi vs. modBi modBi vs. modCi | LRT 6.10398 21.3515 | p-value 0.0473 <.0001 | sign. * | df 8 10 11 | AICc 370.8 369.2 350.2 |
| b) Response varia Model Study design modAi modBi Abundant flowerin modCi Vegetation structu | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area "g ~ width * adj_area + no_flower re | Test modAi vs. modBi modBi vs. modCi | LRT 6.10398 21.3515 | p-value 0.0473 <.0001 | sign. * *** | df 8 10 11 | AICc 370.8 369.2 350.2 |
| b) Response vari Model Study design modAi modBi Abundant flowerin modCi Vegetation structur modDi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area ~ width * adj_area + no_flower re ~ width * adj_area + no_flower + op_ground | Test modAi vs. modBi modBi vs. modCi modCi vs. modDi | LRT 6.10398 21.3515 0.1977 | p-value 0.0473 <.0001 0.6566 | sign. * *** n.s. | df 8 10 11 12 | AICc 370.8 369.2 350.2 352.3 |
| b) Response varia Model Study design modAi modBi Abundant flowerin modCi Vegetation structu modDi modEi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area rg ~ width * adj_area + no_flower re ~ width * adj_area + no_flower + op_ground ~ width * adj_area + no_flower + dom_veghigh | Test modAi vs. modBi modBi vs. modCi modCi vs. modDi modCi vs. modEi | LRT 6.10398 21.3515 0.1977 0.0405 | p-value 0.0473 <.0001 0.6566 0.8404 | sign. * *** n.s. n.s. | df 8 10 11 12 12 | AICc 370.8 369.2 350.2 352.3 352.5 |
| b) Response varia Model Study design modAi modBi Abundant flowerin modCi Vegetation structu modDi modEi Surrounding lands | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area "g ~ width * adj_area + no_flower re ~ width * adj_area + no_flower + op_ground ~ width * adj_area + no_flower + dom_veghigh scope | Test modAi vs. modBi modBi vs. modCi modCi vs. modDi modCi vs. modEi | LRT 6.10398 21.3515 0.1977 0.0405 | p-value 0.0473 <.0001 0.6566 0.8404 | sign. * *** n.s. n.s. | df 8 10 11 12 12 | AICc 370.8 369.2 350.2 352.3 352.5 |
| b) Response varia Model Study design modAi modBi Abundant flowerin modCi Vegetation structu modDi modEi Surrounding lands modFi | able: Total number of butterfly individuals (log(1+indiv)) Variables ~ width + adj_area ~ width * adj_area rg ~ width * adj_area + no_flower re ~ width * adj_area + no_flower + op_ground ~ width * adj_area + no_flower + dom_veghigh scape ~ width * adj_area + no_flower + shan_H | Test modAi vs. modBi modBi vs. modCi modCi vs. modDi modCi vs. modEi modCi vs. modFi | LRT 6.10398 21.3515 0.1977 0.0405 0.0049 | p-value 0.0473 <.0001 0.6566 0.8404 0.9440 | sign. * *** n.s. n.s. n.s. | df 8 10 11 12 12 12 | AICc 370.8 369.2 350.2 352.3 352.5 352.5 |

Signif. codes: ***p < 0.001; **p < 0.01; *p < 0.05; $\cdot p$ < 0.10; n.s. p \geq 0.10. Variables: * = Interaction between the two variables



Fig. 6. Scatterplots of the fixed-effects with significant influence of the number of butterfly species and individuals (Table 6) in field margins (n = 165 observations: 11 inspections per study site, 15 study sites, see Table 1). In order to make the graphic clearly legible, the dots in the scatterplots were jittered.



Fig. 7. Scatterplots of the number of species or individuals in field margins and the interaction between the width of the field margins and the adjacent area (closed squares = field margins next to maize field, open squares = field margins next to flower strips in the first growing season, asterisks = field margins next to flower strips in the second growing season; n = 165 observations: 11 inspections per study site, 15 study sites, see Table 1) In order to make the graphic clearly legible, the dots in the scatterplots were jittered.

species pool in a structurally simple agricultural landscape is limited. Furthermore, species of butterflies that were formerly widespread in the agricultural landscape have also been affected by population declines (Gaston and Fuller, 2007; Haaland et al., 2011; Wallisdevries et al., 2012).

4.2. Factors determining butterflies in flower strips and field margins

4.2.1. Habitat quality

Of all the investigated variables determining the occurrence of butterflies, a high number of flowering plant species was the key driver both in flower strips and in field margins. Interestingly, neither the vegetation height nor the proportion of open-ground had a significant impact on butterflies in either habitats. However, this may be due to the fact that these variables only affect some species and we have considered the total number of species and individuals, or that our data set was too low to detect these effects.

The life span of flower strips can range from several months to several years (Haaland et al., 2011). In our study the flower strips in the first growing season were generally more species-rich (butterflies and flowering plants) and more individual-rich (butterflies) than those in the second growing season. Other studies also showed a reduction in flowering abundance with increasing age of the flower strips, as successional changes may result in an increased proportion of grasses (Frank et al., 2012; Huusela-Veistola and Vasarainen, 2000; Pywell et al., 2011a).

However, a good selection of the seed mixture and a good germination rate and growth can slow down the progress of succession (Aviron et al., 2011). Therefore, the supply of flowering resources over several years must be ensured by a properly selected and planted seed mixture according to the respective soil characteristics (possibly in combination with specific management practices, Haaland et al., 2011).

4.2.2. Landscape context

Our study indicates that flower strips in landscapes with lower habitat diversity (Shannon Diversity Index H) are richer in species and individuals (Table 5). Hence, especially in structurally simple landscapes, which are dominated by arable land that still has a certain proportion of semi-natural habitats (SNH), the flower strips are of high relevance. This result is in line with that of previous studies which have shown that the effect of agri-environmental schemes (AES) is highest in structurally simple, rather than in cleared or in complex landscapes (inter alia for flower strips Scheper et al., 2013, for organic farming: Roschewitz et al., 2005, for local biodiversity conservation management in general Tscharntke et al., 2012). On the other hand, there are studies that have demonstrated the opposite, that biodiversity increases with higher diversity of the surrounding landscape (for field margins Sybertz et al., 2017, for farms Weibull et al., 2000). These conflicting results can be attributed to various causes, such as the differences in study design (survey of different habitats and on different levels (study site level or farm level)), the inconsistent definition of landscape heterogeneity

(Shannon-Index or percentage of different biotopes, varying radii for landscape characteristics) or variations in regional species pools in different landscapes. Above all, differing agricultural intensity of the landscapes seems to be decisive (Ekroos and Kuussaari, 2012; Scheper et al., 2013; Tscharntke et al., 2012). In structurally cleared landscapes (<1% SNH definition after Scheper et al., 2013) as well as in complex landscapes (complex landscapes: > 20% SNH), the effect of AES is low. However, in structurally simple landscapes (1-20% SNH), the SNHs are of great importance for species and their effects are strongest there (Scheper et al., 2013; Tscharntke et al., 2012). We could only detect a significant explanatory power of the habitat diversity of the surrounding landscape in the flower strips and not in the field margins. Obviously, SNH must have a certain quality to produce this positive effect. For a final explanation of the conflicting results concerning the influence of the habitat diversity on the surrounding landscape, more detailed data and analyses are required.

Contrary to our assumption, the connectivity of the landscape had no significant impact on the butterflies in our data set, neither in flower strips nor in field margins. One reason for this could be that the study sites were already sufficiently well connected to the surrounding landscape. Our study area is characterized by a rural environment, where several semi-natural habitats (like grassland, field margins, field paths or hedgerows) are present and provide a network for mobile species such as butterflies. Since for different European butterflies a mean daily displacement from 23-165 m (for sites <0.7 km) or 48–660 m (for sites >0.7 km) could be recorded by different multisite mark-release recapture studies (Stevens et al., 2010).

The analyses of butterfly occurrences on field margins next to flower strips showed that flower strips positively affect adjacent areas: Butterflies were more abundant on field margins next to flower strips than on field margins next to maize fields. Wagner et al. (2014) also demonstrated this effect. They found a higher number of species and individuals on maize fields near flower strips than on maize fields further away from flower strips.

4.3. Conclusions: recommendations to improve the situation of butterflies in the agricultural landscape

Because of their particularly high supply of flowering resources, flower strips are a suitable measure to promote butterflies in the agricultural landscape. Moreover, many other insects can benefit from the creation of flower strips as well (Haaland et al., 2011; Ouvrard et al., 2018; Pywell et al., 2011b). It is crucial to ensure a high number of flowering species by a seed mixture suitable to site conditions and a proper seed bed preparation for germination rate and growth (Aviron et al., 2011).

For butterflies, it is additionally essential that the species composition of the seed mixture takes into account the requirements of the non-adult life stages (Feber et al., 1996; Haaland and Bersier, 2011; Pywell et al., 2011a). The two most frequently detected species in our study (*Artogeia rapae* and *Artogeia napi*) use brassicas as host plants and brassicas were abundant in the "Rotenburger seed mixture 2013" (Supplementary data: Table A.1). Furthermore, the time between sowing and ploughing of the flower strips should not be shorter than two growing seasons, as the non-adult life stages (e.g. the eggs deposited in the vegetation) will be destroyed when the flower strips are converted back into arable land after the first growing season (Haaland and Bersier, 2011).

The optimum age and management of the flower strips must be considered in connection with the seed mixture used. For example, an annual strip needs to be re-sown every year, while legume-dominated pollen and nectar mixtures (as in Pywell et al., 2011a) might last 3–5 years, and perennial wildflower mixtures (as in Aviron et al., 2011) might last up to ten years. Therefore, opinions differ on the optimal age of flower strips. Haaland and Bersier (2011) recommend a minimum life span of five years. Pywell et al. (2011a) showed that flower strips do not offer an optimal number of flowering species after more than three to four years. In contrast, there are results from Switzerland, where up to ten-year-old flower strips without significant influence on butterfly occurrence were found (Aviron et al., 2011). Aviron et al. (2011: 505) stated that 'suitable soil conditions and management of WFS [wildflower strips] allow the maintenance of diversified vegetation over years'. However, Pywell et al. (2011a: 863) explained that flower strips 'are only effective for 3–4 years despite intensive cutting management' and recommended a re-establishment of flower strips after three years in order 'to guarantee a continuity of pollen and nectar resources'. Another reason for the different recommendations could be the differing ecological conditions (i.e. climate, soil) of the study areas in Switzerland (Aviron et al., 2011) and Yorkshire, UK (Pywell et al., 2011a).

Flower strips have the greatest effect in structurally simple landscapes, so they should primarily be used in such landscape sections to counteract biodiversity loss (Scheper et al., 2013; Tscharntke et al., 2012). Flower strips of different growing seasons and with different ploughing periods should be planted next to each other (see also Korpela et al., 2013), because on the one hand, flower strips of different growing seasons have various plant species. On the other hand, the ploughing of the flower strips in an area should not take place in the same year, so that a sufficient supply of flowers is continuously available.

Considerably fewer butterflies were detected in the field margins than in the flower strips. However, the field margins also had a considerably lower number of flowering plant species (Figs. 4 and 6). But on the one hand, the grass-dominated field margins can promote other species than flower strips. In our study, the three most common species in the field margins (Aphantopus hyperantus, Thymelicus lineola and Maniola jurtina) depend on different grass species as host plants. Furthermore, knowing that the amount of flowering resources has a decisive influence on the occurrence of butterflies, the improvement of field margins towards a richer supply of nectar and host plants is particularly interesting. In this way, the field margins could offer a high diversity of flowering resources and also be permanently present in the landscape. So-called 'improved field margins', which are created by sowing a certain mixed seed mixture (annual, perennial herbs and native grasses), are part of the agrienvironmental program in Switzerland (Eggenschwiler et al., 2013; Jacot et al., 2007). They should remain permanently in the landscape and develop into species-rich structural elements with a long-lasting abundance of flowers, for which the annual mowing of half of the strip is recommended. Thus, field margins certainly also have a high potential to counteract biodiversity loss in the agricultural landscape and they must also be considered in conservation measures (Aviron et al., 2011; Kuussaari et al., 2007).

Declarations

Author contribution statement

Nana Wix: Conceived and designed the experiments; Performed the experiments; Analyzed and interpreted the data; Wrote the paper.

Michael Reich: Conceived and designed the experiments; Contributed reagents, materials, analysis tools or data.

Frank Schaarschmid: Contributed reagents, materials, analysis tools or data.

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Competing interest statement

The authors declare no conflict of interest.

Additional information

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References

- Aviron, S., Herzog, F., Klaus, I., Luka, H., Pfiffner, L., Schüpbach, B., 2007a. Effects of Swiss agri-environmental measures on arthropod biodiversity in arable landscapes. Aspect Appl. Biol. 81, 101–109.
- Aviron, S., Herzog, F., Klaus, I., Schüpbach, B., Jeanneret, P., 2011. Effects of wildflower strip quality, quantity, and connectivity on butterfly diversity in a Swiss arable landscape. Restor. Ecol. 19, 500–508.
- Aviron, S., Kindlmann, P., Burel, F., 2007b. Conservation of butterfly populations in dynamic landscapes: the role of farming practices and landscape mosaic. Ecol. Model. 205, 135–145.
- Berg, Å., Ahrné, K., Öckinger, E., Svensson, R., Söderström, B., 2011. Butterfly distribution and abundance is affected by variation in the Swedish forest-farmland landscape. Biol. Conserv. 144, 2819–2831.
- Berg, Å., Ahrné, K., Öckinger, E., Svensson, R., Wissman, J., Stewart, A., Bezemer, M., 2013. Butterflies in semi-natural pastures and power-line corridors - effects of flower richness, management, and structural vegetation characteristics. Insect Conserv Divers 6, 639–657.
- Bolker, B.M., Brooks, M.E., Clark, C.J., Geange, S.W., Poulsen, J.R., Stevens, M.H.H., White, J.-S.S., 2009. Generalized linear mixed models: a practical guide for ecology and evolution. Trends Ecol. Evol. 24, 127–135.
- Brittain, C.A., Vighi, M., Bommarco, R., Settele, J., Potts, S.G., 2010. Impacts of a pesticide on pollinator species richness at different spatial scales. Basic Appl. Ecol. 11, 106–115.
- Brückmann, S.V., Krauss, J., Steffan-Dewenter, I., 2010. Butterfly and plant specialists suffer from reduced connectivity in fragmented landscapes. J. Appl. Ecol. 47, 799–809.
- Burnham, K.P., Anderson, D.R., 2002. Model Selection and Multimodel Inference: A Practical Information-Theoretic Approach. Springer, New York, p. 208.
- Clausen, H.D., Holbeck, H.B., Reddersen, J., 2001. Factors influencing abundance of butterflies and burnet moths in the uncultivated habitats of an organic farm in Denmark. Biol. Conserv. 98, 167–178.
- Cole, L.J., Brocklehurst, S., Robertson, D., Harrison, W., McCracken, D.I., 2015. Riparian buffer strips: their role in the conservation of insect pollinators in intensive grassland systems. Agric. Ecosyst. Environ. 211, 207–220.
- Dainese, M., Luna, D.I., Sitzia, T., Marini, L., 2015. Testing scale-dependent effects of seminatural habitats on farmland biodiversity. Ecol. Appl. 25, 1681–1690.
- Delattre, T., Pichancourt, J.-B., Burel, F., Kindlmann, P., 2010. Grassy field margins as potential corridors for butterflies in agricultural landscapes: a simulation study. Ecol. Model. 221, 370–377.
- Delattre, T., Vernon, P., Burel, F., 2013. An agri-environmental scheme enhances butterfly dispersal in European agricultural landscapes. Agric. Ecosyst. Environ. 166, 102–109. Dollar, J.G., Riffell, S.K., Burger, L.W., 2013. Effects of managing semi-natural grassland
- buffers on butterfly distribution Dover, J., Settele, J., 2009. The influences of landscape structure on butterfly distribution
- and movement: a review. J. Insect Conserv. 13, 3–27. Dover, J., Sparks, T., Clarke, S., Gobbett, K., Glossop, S., 2000. Linear features and
- butterflies: the importance of green lanes. Agric. Ecosyst. Environ. 80, 227–242.
- Dover, J.W., 1996. Factors affecting the distribution of satyrid butterflies on arable farmland. J. Appl. Ecol. 33, 723.
- Dover, J.W., Sparks, T.H., Greatorex-Davis, F.N., 1997. The importance of shelter for butterflies in open landscapes. J. Insect Conserv. 1, 89–97.
- Eggenschwiler, L., Speiser, B., Bosshard, A., Jacot, K., 2013. Improved field margins highly increase slug activity in Switzerland. Agron. Sustain. Dev. 33, 349–354.
- Ekroos, J., Kuussaari, M., 2012. Landscape context affects the relationship between local and landscape species richness of butterflies in semi-natural habitats. Ecography 35, 232–238.
- Ekroos, J., Kuussaari, M., Tiainen, J., Heliölä, J., Seimola, T., Helenius, J., 2013. Correlations in species richness between taxa depend on habitat, scale and landscape context. Ecol. Indicat. 34, 528–535.
- Ekroos, J., Piha, M., Tiainen, J., 2008. Role of organic and conventional field boundaries on boreal bumblebees and butterflies. Agric. Ecosyst. Environ. 124, 155–159.
- Feber, R.E., Smith, H., Macdonald, D.W., 1996. The effects on butterfly abundance of the management of uncropped edges of arable fields. J. Appl. Ecol. 33, 1191–1205.

- Field, R.G., Gardiner, T., Mason, C.F., Hill, J., 2005. Agri-environment schemes and butterflies: the utilisation of 6 m grass margins. Biodivers. Conserv. 14, 1969–1976.
- Field, R.G., Gardiner, T., Mason, C.F., Hill, J., 2006. Countryside stewardship scheme and butterflies: a study of plant and butterfly species richness. Biodivers. Conserv. 15, 443–452.
- Field, R.G., Gardiner, T., Mason, C.F., Hill, J., 2007. Agri-environment schemes and butterflies: the utilisation of two metre arable field margins. Biodivers. Conserv. 16, 465–474.
- Fox, R., Brereton, T.M., Asher, J., August, T.A., Botham, M.S., Bourn, N.A.D., Cruickshanks, K.L., Bulman, C.R., Ellis, S., Harrower, C.A., Middlebrook, I., Noble, D.G., Powney, G.D., Randle, Z., Warren, M.S., Roy, D.B., 2015. The State of the UK's Butterflies 2015. In: Butterfly Conservation and the Centre for Ecology & Hydrology. Wareham, Dorset.
- Frank, T., Aeschbacher, S., Zaller, J.G., 2012. Habitat age affects beetle diversity in wildflower areas. Agric. Ecosyst. Environ. 152, 21–26.
- Gabriel, D., Sait, S.M., Hodgson, J.A., Schmutz, U., Kunin, W.E., Benton, T.G., 2010. Scale matters: the impact of organic farming on biodiversity at different spatial scales. Ecol. Lett. 13, 858–869.
- Gaston, K.J., Fuller, R.A., 2007. Biodiversity and extinction: losing the common and the widespread. Prog. Phys. Geogr. 31, 213–225.
- GeoBasis-DE/BKG, 2017. Verwaltungsgebiete 1:2 500 000 Stand 01.01.2017. http ://www.geodatenzentrum.de/geodaten/gdz_rahmen.gdz_div?gdz_spr=deu&gdz_ akt_zeile=5&gdz_anz_zeile=1&gdz_unt_zeile=19&gdz_user_id=0. (Accessed 23 January 2018).
- GeoBasis-DE/BKG, 2012. CORINE Land Cover 10 Ha.
- http://www.geodatenzentrum.de/geodaten/gdz_rahmen.gdz_div?gdz_spr=deu& gdz_akt_zeile=5&gdz_anz_zeile=1&gdz_unt_zeile=22&gdz_user_id=0#dok. (Accessed 24 January 2018).
- Giuliano, D., Cardarelli, E., Bogliani, G., 2018. Grass management intensity affects butterfly and orthopteran diversity on rice field banks. Agric. Ecosyst. Environ. 267, 147–155.
- Haaland, C., Bersier, L.-F., 2011. What can sown wildflower strips contribute to butterfly conservation?: an example from a Swiss lowland agricultural landscape. J. Insect Conserv. 15, 301–309.
- Haaland, C., Gyllin, M., 2010. Butterflies and bumblebees in greenways and sown wildflower strips in southern Sweden. J. Insect Conserv. 14, 125–132.
- Haaland, C., Naisbit, R.E., Bersier, L.-F., 2011. Sown wildflower strips for insect conservation: a review. Insect Conservation and Diversity 4, 60–80.
- Hambler, C., Henderson, P.A., Speight, M.R., 2011. Extinction rates, extinction-prone habitats, and indicator groups in Britain and at larger scales. Biol. Conserv. 144, 713–721.
- Holland, J.M., Smith, B.M., Storkey, J., Lutman, P.J.W., Aebischer, N.J., 2015. Managing habitats on English farmland for insect pollinator conservation. Biol. Conserv. 182, 215–222.
- Huusela-Veistola, E., Vasarainen, A., 2000. Plant succession in perennial grass strips and effects on the diversity of leafhoppers (Homoptera, Auchenorrhyncha). Agric. Ecosyst. Environ. 80, 101–112.
- Jacot, K., Eggenschwiler, L., Junge, X., Luka, H., Bosshard, A., 2007. Improved field margins for a higher biodiversity in agricultural landscapes. Aspect Appl. Biol.
- Jaeger, J.A.G., 2000. Landscape division, splitting index, and effective mesh size: new measures of landscape fragmentation. Landsc. Ecol. 15, 115–130.
- Jonason, D., Andersson, G.K.S., Ockinger, E., Rundlöf, M., Smith, H.G., Bengtsson, J., 2011. Assessing the effect of the time since transition to organic farming on plants and butterflies. J. Appl. Ecol. 48, 543–550.
- Kamil, B., 2018. MuMIn: Multi-Model Inference. R Package Version 1.42.1. https:// CRAN.R-project.org/package=MuMIn.
- Korpela, E.-L., Hyvönen, T., Lindgren, S., Kuussaari, M., 2013. Can pollination services, species diversity and conservation be simultaneously promoted by sown wildflower strips on farmland? Agric. Ecosyst. Environ. 179, 18–24.
- Kruse, M., Stein-Bachinger, K., Gottwald, F., Schmidt, E., Heinken, T., 2016. Influence of grassland management on the biodiversity of plants and butterflies on organic suckler cow farms. Tuexenia 97–119.
- Kuckartz, U., R\u00e4diker, S., Ebert, T., Schehl, J., 2013. Statistik: Eine Verst\u00e4ndliche Einf\u00fchrung, second ed. Springer VS, Wiesbaden, p. 39.
- Kuussaari, M., Heliölä, J., Luoto, M., Pöyry, J., 2007. Determinants of local species richness of diurnal Lepidoptera in boreal agricultural landscapes. Agric. Ecosyst. Environ. 122, 366–376.
- Lenth, R., 2018. Emmeans: Estimated Marginal Means, Aka Least-Squares Means. R Package Version 1.3.0. https://CRAN.R-project.org/package=emmeans.
- LSN (Landesamt für Statistik Niedersachsen), 2018. Bodenflächen in Niedersachsen nach Art der tatsächlichen Nutzung 2016 – Stand: 31.12.2015 –. Statistische Berichte Niedersachsen. http://www.statistik.niedersachsen.de/themenbereiche/flaechenn utzung_gebietsstand/themenbereich-flaechennutzung-gebietsstand—statistische-be richte-87671.html. (Accessed 6 September 2018).
- Lang, S., Blaschke, T., 2007. Landschaftsanalyse mit GIS: 20 Tabellen, first ed. Ulmer, Stuttgart, p. 405.
- Lebeau, J., Wesselingh, R.A., Dyck, H., 2016. Floral resource limitation severely reduces butterfly survival, condition and flight activity in simplified agricultural landscapes. Oecologia 180, 421–427.
- Luppi, M., Dondina, O., Orioli, V., Bani, L., 2018. Local and landscape drivers of butterfly richness and abundance in a human-dominated area. Agric. Ecosyst. Environ. 254, 138–148.
- Meek, B., Loxton, D., Sparks, T., Pywell, R., Pickett, H., Nowakowski, M., 2002. The effect of arable field margin composition on invertebrate biodiversity. Biol. Conserv. 106, 259–271.

MU Nds, 2018. (Niedersächsisches Ministerium für Umwelt, Energie, Bauen und Klimaschutz). URI-Liste für WMS-Dienste des Kartenservers des MU. Basisdaten. htt ps://www.umwelt.niedersachsen.de/service/umweltkarten/wmsdienste/url-listefuer-wms-dienste-des-kartenservers-des-mu-8887.html. (Accessed 23 January 2018).

- Noordijk, J., Delille, K., Schaffers, A.P., Sýkora, K.V., 2009. Optimizing grassland management for flower-visiting insects in roadside verges. Biol. Conserv. 142, 2097–2103.
- Ouin, A., Burel, F., 2002. Influence of herbaceous elements on butterfly diversity in hedgerow agricultural landscapes. Agric. Ecosyst. Environ. 93, 45–53.
- Ouvrard, P., Transon, J., Jacquemart, A.-L., 2018. Flower-strip agri-environment schemes provide diverse and valuable summer flower resources for pollinating insects. Biodivers. Conserv. 27, 2193–2216.
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., R Core Team, 2018. Nlme: Linear and Nonlinear Mixed Effects Models. R Package Version 3. https://CRAN.R-project.org/ package=nlme. (Accessed 24 September 2018).
- Potts, S.G., Biesmeijer, J.C., Kremen, C., Neumann, P., Schweiger, O., Kunin, W.E., 2010. Global pollinator declines: trends, impacts and drivers. Trends Ecol. Evol. 25, 345–353.
- Pywell, R.F., Meek, W.R., Hulmes, L., Hulmes, S., James, K.L., Nowakowski, M., Carvell, C., 2011a. Management to enhance pollen and nectar resources for bumblebees and butterflies within intensively farmed landscapes. J. Insect Conserv. 15, 853–864.
- Pywell, R.F., Meek, W.R., Loxton, R.G., Nowakowski, M., Carvell, C., Woodcock, B.A., 2011b. Ecological restoration on farmland can drive beneficial functional responses in plant and invertebrate communities. Agric. Ecosyst. Environ. 140, 62–67.
- Pywell, R.F., Warman, E.A., Sparks, T.H., Greatorex-Davies, J.N., Walker, K.J., Meek, W.R., Carvell, C., Petit, S., Firbank, L.G., 2004. Assessing habitat quality for butterflies on intensively managed arable farmland. Biol. Conserv. 118, 313–325.
- Quinn, G.P., Keough, M.J., 2014. Experimental Design and Data Analysis for Biologists, twelfth ed. Cambridge Univ. Press, Cambridge, p. 537.
- Reinhardt, R., Bolz, R., 2011. Rote Liste und Gesamtartenliste der Tagfalter (Rhopalocera) (Lepidoptera: papilionoidea et Hesperoidea) Deutschlands. Natursch. Biol. Vielfalt 70, 167–194.
- Roschewitz, I., Gabriel, D., Tscharntke, T., Thies, C., 2005. The effects of landscape complexity on arable weed species diversity in organic and conventional farming. J. Appl. Ecol. 42, 873–882.
- RStudio Team, 2016. RStudio. Integrated Development for R. R Studio, Inc., Boston.
- Rundlöf, M., Bengtsson, J., Smith, H.G., 2008. Local and landscape effects of organic farming on butterfly species richness and abundance. J. Appl. Ecol. 45, 813–820.
- Rundlöf, M., Smith, H.G., 2006. The effect of organic farming on butterfly diversity depends on landscape context. J. Appl. Ecol. 43, 1121–1127.
- Saarinen, K., 2002. A comparison of butterfly communities along field margins under traditional and intensive management in SE Finland. Agric. Ecosyst. Environ. 90, 59–65.
- Saarinen, K., Valtonen, A., Jantunen, J., Saarnio, S., 2005. Butterflies and diurnal moths along road verges: does road type affect diversity and abundance? Biol. Conserv. 123, 403–412.
- Šálek, M., Hula, V., Kipson, M., Daňková, R., Niedobová, J., Gamero, A., 2018. Bringing diversity back to agriculture: smaller fields and non-crop elements enhance biodiversity in intensively managed arable farmlands. Ecol. Indicat. 90, 65–73.
- Scheper, J., Holzschuh, A., Kuussaari, M., Potts, S.G., Rundlöf, M., Smith, H.G., Kleijn, D., Gomez, J., 2013. Environmental factors driving the effectiveness of European agrienvironmental measures in mitigating pollinator loss - a meta-analysis. Ecol. Lett. 16, 912–920.
- Settele, J. (Ed.), 1999. Die Tagfalter Deutschlands: 48 Tabellen. Ulmer, Stuttgart, p. 452. Settele, J., Steiner, R., 2009. Schmetterlinge: Die Tagfalter Deutschlands. Ulmer, Stuttgart (Hohenheim), p. 256.
- Skórka, P., Lenda, M., Moroń, D., Kalarus, K., Tryjanowski, P., 2013. Factors affecting road mortality and the suitability of road verges for butterflies. Biol. Conserv. 159, 148–157.
- Snoo, G.R., Poll, R.J., Bertels, J., 1998. Butterflies in sprayed and unsprayed field margins. J. Appl. Entomol. 122, 157–161.
- Snoo, G.R., 1999. Unsprayed field margins: effects on environment, biodiversity and agricultural practice. Landsc. Urban Plann. 46, 151–160.
- Sparks, T.H., Parish, T., 1995. Factors affecting the abundance of butterflies in field boundaries in Swavesey fens, Cambridgeshire, UK. Biol. Conserv. 73, 221–227.

- Stevens, V.M., Turlure, C., Baguette, M., 2010. A meta-analysis of dispersal in butterflies. Biol. Rev. Camb. Philos. Soc. 85, 625–642.
- Sutter, L., Albrecht, M., Jeanneret, P., 2018. Landscape greening and local creation of wildflower strips and hedgerows promote multiple ecosystem services. J. Appl. Ecol. 55, 612–620.
- Swaay, C.A.M., Warren, M., Loïs, G., 2006. Biotope use and trends of european butterflies. J. Insect Conserv. 10, 189–209.
- Sybertz, J., Matthies, S., Schaarschmidt, F., Reich, M., Haaren, C., 2017. Assessing the value of field margins for butterflies and plants: how to document and enhance biodiversity at the farm scale. Agric. Ecosyst. Environ. 249, 165–176.
- Taylor, M.E., Morecroft, M.D., 2009. Effects of agri-environment schemes in a long-term ecological time series. Agric. Ecosyst. Environ. 130, 9–15.
- Thomas, J.A., Telfer, M.G., Roy, D.B., Preston, C.D., Greenwood, J.J.D., Asher, J., Fox, R., Clarke, R.T., Lawton, J.H., 2004. Comparative losses of British butterflies, birds, and plants and the global extinction crisis. Science (New York, N.Y.) 303, 1879–1881.
- Toivonen, M., Peltonen, A., Herzon, I., Heliölä, J., Leikola, N., Kuussaari, M., Didham, R., Batary, P., 2017. High cover of forest increases the abundance of most grassland butterflies in boreal farmland. Insect Conserv Divers 10, 321–330.
- Toivonen, M., Herzon, I., Kuussaari, M., 2016. Community composition of butterflies and bumblebees in fallows: niche breadth and dispersal capacity modify responses to fallow type and landscape. J. Insect Conserv. 20, 23–34.
- Toivonen, M., Herzon, I., Kuussaari, M., 2015. Differing effects of fallow type and landscape structure on the occurrence of plants, pollinators and birds on environmental fallows in Finland. Biol. Conserv. 181, 36–43.
- Tolman, T., Lewington, R., 2009. Collins Butterfly Guide: the Most Complete Guide to the Butterflies of Britain and Europe. Collins, London, p. 384.
- Traxler, A., 1997. Handbuch des Vegetationsökologischen Monitorings: Methoden, Praxis, angewandte Projekte, Teil A. Methoden. Monographien 89A, Wien.
- Tscharntke, T., Tylianakis, J.M., Rand, T.A., Didham, R.K., Fahrig, L., Batáry, P., Bengtsson, J., Clough, Y., Crist, T.O., Dormann, C.F., Ewers, R.M., Fründ, J., Holt, R.D., Holzschuh, A., Klein, A.M., Kleijn, D., Kremen, C., Landis, D.A., Laurance, W., Lindenmayer, D., Scherber, C., Sodhi, N., Steffan-Dewenter, I., Thies, C., der Putten, W.H., Westphal, C., 2012. Landscape moderation of biodiversity patterns and processes - eight hypotheses. Biol. Rev. Camb. Philos. Soc. 87, 661–685.
- Uyttenbroeck, R., Hatt, S., Piqueray, J., Paul, A., Bodson, B., Francis, F., Monty, A., 2015. Creating perennial flower strips: think functional! Agriculture and Agricultural Science Procedia 6, 95–101.
- Uyttenbroeck, R., Piqueray, J., Hatt, S., Mahy, G., Monty, A., 2017. Increasing plant functional diversity is not the key for supporting pollinators in wildflower strips. Agric. Ecosyst. Environ. 249, 144–155.
- Valtonen, A., Saarinen, K., Jantunen, J., 2006. Effect of different mowing regimes on butterflies and diurnal moths on road verges. Anim. Biodivers. Conserv. 29, 133–148.
- Wagner, C., Bachl-Staudinger, M., Baumholzer, S., Burmeister, J., Fischer, C., Karl, N., Köppl, A., Volz, H., Walter, R., Wieland, P., 2014. Faunistische Evaluierung von Blühflächen. Schriftenreihe der Bayerischen Landesanstalt für Landwirtschaft, 1–150. https://www.lfl.bayern.de/publikationen/schriftenreihe/059344/index.php.
- Wallisdevries, M.F., Swaay, C.A.M., Plate, C.L., 2012. Changes in nectar supply: a possible cause of widespread butterfly decline. Current Zoology 58, 384.
- Warren, M.S., Hill, J.K., Thomas, J.A., Asher, J., Fox, R., Huntley, B., Roy, D.B., Telfer, M.G., Jeffcoate, S., Harding, P., Jeffcoate, G., Willis, S.G., Greatorex-Davies, J.N., Moss, D., Thomas, C.D., 2001. Rapid responses of British butterflies to opposing forces of climate and habitat change. Nature 414, 65.
- Weibull, A., Ostman, O., Granqvist, A., 2003. Species richness in agroecosystems: the effect of landscape, habitat and farm management. Biodivers. Conserv. 12, 1335–1355.
- Weibull, A.-C., Bengtsson, J., Nohlgren, E., 2000. Diversity of butterflies in the agricultural landscape: the role of farming system and landscape heterogeneity. Ecography 23, 743–750.
- Wickham, H., 2016. ggplot2: Elegant Graphics for Data Analysis. Springer-Verlag, New York.
- Woodcock, B.A., Savage, J., Bullock, J.M., Nowakowski, M., Orr, R., Tallowin, J.R.B., Pywell, R.F., 2014. Enhancing floral resources for pollinators in productive agricultural grasslands. Biol. Conserv. 171, 44–51.
- Wix, N., Reich, M., 2018. Die Tagfalterfauna von Blühstreifen. Umwelt und Raum 9, 223–253.
- Zuur, A.F., Ieno, E.N., Walker, N., Saveliev, A.A., Smith, G.M., 2009. Mixed Effects Models and Extensions in Ecology with R. Springer, New York, NY.